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添加微塑料和水稻秸秆对典型农田土壤 CO₂ 排放的影响

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摘要 为探究微塑料污染对水稻秸秆还田下土壤有机碳矿化的影响, 采集3种典型农田土壤(水稻土、红壤和潮土), 进行70 d室内培养试验, 每种土壤设置对照(control check, CK)、添加微塑料(polyethylene, PE)、添加水稻秸秆(rice straw, RS)、添加微塑料和水稻秸秆(rice straw-polyethylene, RS-PE)4个处理, 测定各处理土壤理化性质、CO₂ 释放量、可溶性有机碳(soluble organic carbon, DOC)和微生物生物量碳(microbial biomass carbon, MBC)含量。结果显示: 3种土壤中PE和RS-PE处理下70 d内CO₂ 累积释放量大小为潮土> 稻土> 红壤。与CK处理相比, RS处理后3种土壤CO₂ 累积释放量均显著增加, PE处理后红壤和潮土中CO₂ 累积释放量显著增加。添加微塑料和水稻秸秆提高了3种土壤DOC和MBC含量, 促进土壤有机碳矿化。3种土壤中平均相对分子质量(E₂/E₃)值依次为水稻土> 红壤> 潮土; 且添加微塑料和水稻秸秆提高了土壤DOC平均相对分子质量、芳香性和疏水性。3种土壤有机碳矿化均与DOC E₂/E₃呈显著负相关, 且水稻土和潮土有机碳矿化与DOC芳香性(SUVA₂₅₄)和疏水性(SUVA₂₆₀)呈显著正相关。研究结果表明, 微塑料和水稻秸秆还田显著影响土壤有机碳的矿化。

关键词 微塑料; 土壤; 水稻秸秆; 有机碳矿化; CO₂ 排放

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土壤有机碳是陆地生态系统最大的碳库^[1], 其矿化是土壤中重要的生物地球化学过程^[2-3], 与土壤有机碳稳定和温室气体排放密切相关。开展土壤固碳减排研究, 对调控温室效应和实现“双碳”目标具有重要意义。

据欧洲塑料协会统计, 2020年全球塑料产量达到3.67亿t, 且预测仍将呈上涨趋势^[4]。大部分废弃塑料进入环境后经过光照、高温氧化、物理风化和生物降解等作用形成粒径小于5 mm的塑料颗粒或碎片, 即微塑料(microplastics, MPs)^[5-6]。土壤中的MPs来源广泛, 与人类活动关系密切, 主要包括塑料废弃物、地膜覆盖、污泥堆肥、污灌和大气沉降等^[7-8]。MPs进入土壤后改变土壤结构和通气透水能力^[9-10], 且进入土壤孔隙的MPs可能被土壤微生物作为潜在的碳源, 进而影响碳转化相关微生物的

数量和群落^[11-12], 最终可能影响土壤有机碳组分和温室气体如CO₂的排放。

水稻秸秆还田作为资源高效和可持续利用方式, 能有效改善土壤环境、增加土壤固碳量及土壤养分, 降低稻区温室效应和温室气体排放^[13]。Wang等^[14]的Meta分析结果显示, 秸秆还田可使土壤有机碳含量增加13.97%。与不还田相比, 秸秆还田显著提高土壤中可溶性有机碳和微生物量碳含量^[15], 进而影响有机碳分解与矿化过程。微塑料与水稻秸秆进入土壤后, 可能对土壤碳循环产生一定影响。基于此, 本研究将微塑料污染与秸秆还田相结合, 通过对3种典型农田土壤添加聚乙烯微塑料和水稻秸秆, 探究微塑料对秸秆添加下土壤有机碳转化及CO₂排放特征, 以期明确微塑料污染和秸秆还田对土壤碳排放影响提供科学依据。

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1 材料与方法

1.1 供试材料

选取水稻土、红壤和潮土为供试土壤,分别取自湖北省荆门市(30° 49'N、112° 9'E)、咸宁市(30° 01'N, 114° 21'E)和武汉市(30° 38'N、114° 15'E)。采集耕作层(0~20 cm)土壤,去除可见的砾石和动植物残体,运回实验室后通风阴干。一部分土壤过孔径 2 mm 和 0.149 mm 筛,用于土壤理

化性质(表1)的测定,剩余土壤过过孔径 2 mm 筛用于培养试验。

供试聚乙烯微塑料(PE)购自中诚塑化有限公司,粒径小于 0.149 mm。供试水稻秸秆(RS)在湖北省武汉市华中农业大学试验基地水稻成熟期采集,风干后粉碎,过 0.25 mm 筛备用。使用 pH 计和元素分析仪测定 3 种土壤、聚乙烯微塑料和水稻秸秆的基本性质,结果如表 1 所示。

表 1 供试材料的基本性质

Table 1 Basic properties of tested materials

材料 Materials	总有机碳/(g/kg) Total organic carbon	全氮/(g/kg) Total nitrogen	碳氮比 C/N	pH
水稻土 Paddy soil	19.51±0.05	1.79±0.08	12.71	6.07±0.10
红壤 Red soil	11.69±0.07	1.16±0.12	11.76	4.14±0.21
潮土 Fluvo-aquic soil	7.23±0.14	0.73±0.05	11.55	7.58±0.11
水稻秸秆 Rice straw	356.48±0.55	11.38±0.14	36.55	
聚乙烯塑料 Polyethylene	808.35±0.83	0.08±0.01		

1.2 培养试验

各土壤样品在恒温恒湿、黑暗条件下预培养 7 d,以激活土壤微生物活性。3 种土壤(水稻土、红壤和潮土)分别设置 4 个处理:对照组(不添加任何材料,CK);微塑料组(添加 0.5% 微塑料,PE);秸秆组(添加 0.25% 水稻秸秆,RS);水稻秸秆+塑料组(添加 0.25% 水稻秸秆和 0.5% 微塑料,RS-PE),每个处理设 3 次重复。

称取预培养土壤 60 g 于 500 mL 培养瓶中,分别按上述处理添加试验材料,调节含水量为田间含水量(WHC)的 60%,培养瓶口用带小孔的保鲜膜密封,25 °C 黑暗条件下好气培养 70 d,按称质量法定期补充水分。分别在培养的第 1、3、7、15、30、50、70 天时采集气体,用于测定 CO₂ 浓度。培养结束后将培养瓶内样品混匀、风干、过筛,用于其他理化性质分析。

1.3 测定指标及方法

1) 基本理化性质分析。称取过孔径 2 mm 筛的土样于 50 mL 带盖离心管,按照土水比 1:2.5(m/V)向离心管中加入去 CO₂ 的蒸馏水,震荡后静置 30 min,用 pH 计(FiveEasy Plus FE28)测定悬液的 pH 值^[16]。称取一定量土壤,用元素分析仪(Isoprime 100)测定各土壤全氮、全碳含量,计算碳氮比(C/N)。

2) CO₂ 释放量的测定。采集的气体用气相色谱仪(Agilent, G8890A)测定 CO₂ 浓度。

3) 可溶性有机碳和微生物量碳的测定。土壤可溶性有机碳(soluble organic carbon, DOC)含量和结构的测定参照 Xu 等^[17]的方法。土样按 1:5(m/V)的

比例加入超纯水,室温下振荡 1 h 后以 10 000 r/min 离心 5 min,悬液过 0.45 μm 滤膜。用总有机碳分析仪(德国 Elementer Vario)测定浸提液有机碳浓度,计算 DOC 含量。

用紫外-可见分光光度计(UV-1500,日本 Shimadzu)测定 DOC 的紫外吸收光谱。计算 254 nm(SUVA₂₅₄)和 260 nm(SUVA₂₆₀)处的紫外吸光度,分别表示 DOC 的芳香度和疏水性。250 nm(E₂)和 365 nm(E₃)处吸光度的比值表示 DOC 的平均相对分子质量大小^[18]。

使用荧光光谱仪(F-7100,日本 Hitachi)测定 DOC 的三维荧光光谱(fluorescence excitation emission matrix, EEM)^[17]。

微生物生物量碳(microbial biomass carbon, MBC)含量用氯仿熏蒸-K₂SO₄ 浸提法^[16],熏蒸与未熏蒸样品 DOC 含量之差,除以转换系数(K_c=0.45)而得。

1.4 数据统计分析

根据 Murphy 等^[19]的方法,采用 MATLAB(MathWorks,美国 Natick)中的 drEEM 0.6.0 工具箱进行平行因子分析(PARAFAC)。经过拉曼归一化、去拉曼散射和瑞利散射^[20]、残差分析、半分检验^[21]后,最终经反复迭代确定 DOC 的最佳组分为三组分模型(命名为 C1、C2、C3)。最佳组分的相对含量根据 PARAFAC 获得的最大荧光强度(F_{max})得出。

数据统计及方差分析采用 SPSS 2021 软件完

成,用Microsoft Excel 2021和Origin 2021软件进行数据处理并绘图。双因素方差分析结合Duncan's检验比较土壤类型和外源材料添加对测试指标影响的差异。用Pearson相关评估指标之间的相关性。

2 结果与分析

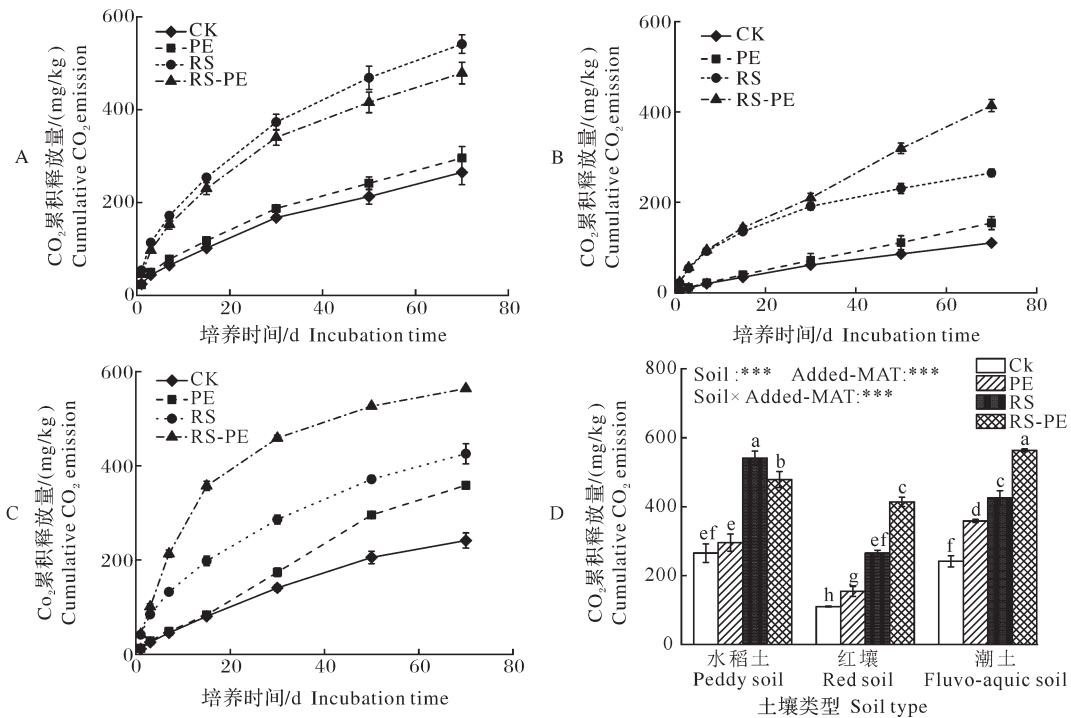
2.1 微塑料和秸秆添加对土壤二氧化碳排放量的影响

在70 d培养期间,水稻土和潮土中CO₂累积释放量呈前期(0~30 d)迅速增加、后期(30~70 d)增加较缓的趋势(图1 A、C);而在红壤中PE和RS-PE处理CO₂累积释放量在整个培养期间快速增加(图1 B)。在70 d培养期内,水稻土CK、PE、RS、RS-PE处理中CO₂累积排放量分别为265.2、295.8、541.1、478.7 mg/kg,单独秸秆施用使水稻土中CO₂累积排放量增加了104%,微塑料在不施用秸秆时使水稻土中CO₂累积排放量增加了12%,但在施用秸秆时使水稻土中CO₂累积排放量减少了12%。

在70 d培养期内,红壤CK、PE、RS、RS-PE处理中CO₂累积排放量分别为110.3、154.2、265.3、414.0

mg/kg,单独秸秆施用使红壤中CO₂累积排放量增加141%,微塑料在不施用和施用秸秆时使红壤中CO₂累积排放量分别增加了40%和56%。在70 d的培养期内,潮土CK、PE、RS、RS-PE处理中CO₂累积排放量分别为241.6、358.5、425.5、563.6 mg/kg,单独秸秆施用使潮土中CO₂累积排放量增加了76%,微塑料在不施用和施用秸秆时使潮土中CO₂累积排放量分别增加了48%和32%。

3种土壤中PE和RS-PE处理下70 d的CO₂累积释放量为潮土>水稻土>红壤(图1 D)。土壤类型、添加材料和两者交互作用对70 d内CO₂累积释放量影响显著(图1 D)。单独秸秆施用增加了供试3种土壤的CO₂累积排放量,但增幅以红壤中最高,而潮土中最低;微塑料在不施用秸秆时也均增加供试3种土壤的CO₂累积排放量,但增幅均远小于相应单独秸秆施用的处理,且增幅在水稻土中最低;微塑料在施用秸秆时增加了红壤和潮土中的CO₂累积排放量,但减少了水稻土中的CO₂累积排放量,这可能与水稻土的固碳能力强有关。



A:水稻土;B:红壤;C:潮土;D:累积释放量。双因素方差分析比较土壤类型(Soil)和外源材料添加(Added-MAT)对测试指标影响的差异,***表示影响因子的差异极显著, $P < 0.001$ 。不同小写字母表示差异显著($P < 0.05$);下同。A:Paddy soil;B:Red soil;C:Fluvo-aquic soil;;D: CO₂-C cumulative emissions.Two-way ANOVA comparing the differences for soil type (Soil) and exogenous materials addition (Added-MAT) on the test metrics, *** indicating significant differences for the impact factors, $P < 0.001$.Different letters indicate significant differences ($P < 0.05$). The same as below.

图1 培养过程中CO₂-C释放动态与累积释放量

Fig 1 Cumulative CO₂-C emission during co-incubation with exogenously added materials during 70 days

2.2 微塑料和秸秆添加对土壤理化性质的影响

在3种土壤中,与CK相比,除红壤添加秸秆处理外,微塑料和秸秆添加均显著增加TOC含量,且土壤类型和外源材料添加及其交互作用对TOC的影响显著(表2)。无论是否添加水稻秸秆,在3种土壤中添加微塑料均增加土壤碳氮比(C/N),且土壤类型和外源材料添加的交互作用对C/N影响极显著($P < 0.001$)。土壤pH值高低为潮土>水稻土>红壤,与外源材料添加无关。

2.3 微塑料和秸秆添加对土壤DOC和MBC含量的影响

3种土壤同一处理下可溶性有机碳(DOC)和微

生物量碳(MBC)含量由高到低为水稻土>红壤>潮土(图2),且MBC含量差异较大。水稻土、红壤和潮土中添加水稻秸秆后,DOC含量相比对照分别增加19.11、8.88和4.51 mg/kg,均大于添加微塑料比对照的增加量4.04、2.76和2.55 mg/kg。水稻土不加秸秆时,微塑料的加入使土壤MBC含量增加12.15%;添加秸秆后,微塑料对土壤MBC含量影响不显著。红壤不添加秸秆时,微塑料的加入使土壤MBC含量增加38.03%;添加秸秆后,微塑料对土壤MBC含量影响不显著。潮土在未添加秸秆时,微塑料的加入使土壤MBC含量增加17.69%;添加秸秆后,微塑料的加入使土壤MBC含量增加30.88%。

表2 土壤与微塑料、水稻秸秆及二者共培养后的基本性质

Table 2 The properties of soils co-incubation with microplastics, rice straw and both

土壤类型 Soil type	处理 Treatment	TOC(g/kg)	TN/(g/kg)	C/N	pH
水稻土 Paddy soil	CK	20.41±0.08c	2.15±0.02b	11.08	5.96±0.07b
	PE	23.74±0.10a	2.07±0.02c	13.38	6.04±0.06b
	RS	21.30±0.09b	2.21±0.02a	11.24	6.04±0.04b
	RS-PE	24.05±0.35a	2.10±0.01c	13.36	6.09±0.02b
红壤 Red soil	CK	12.83±0.02e	1.42±0.01e	10.54	4.12±0.08c
	PE	15.29±0.04d	1.41±0.02e	12.65	4.15±0.04c
	RS	12.78±0.36e	1.47±0.03d	10.14	4.12±0.09c
	RS-PE	15.25±0.16d	1.40±0.01e	12.71	4.22±0.04c
潮土 Fluvo-aquic soil	CK	7.97±0.06i	0.95±0.02h	9.79	7.61±0.05a
	PE	10.69±0.13g	1.00±0.02g	12.47	7.50±0.03a
	RS	8.61±0.18h	1.06±0.02f	9.47	7.53±0.07a
	RS-PE	11.27±0.01f	0.99±0.04gh	13.28	7.62±0.10a

注:同一列中不同小写字母表示差异显著($P < 0.05$)。Note: Different lowercase letters in the same column indicate significant differences ($P < 0.05$).

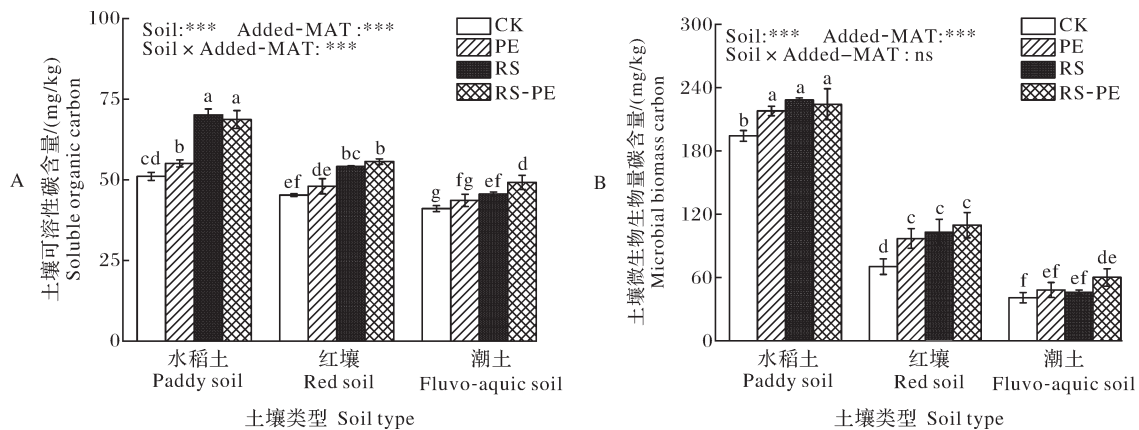


图2 土壤与微塑料、水稻秸秆共培养后可溶性有机碳(A)和微生物量碳(B)含量

Fig. 2 Soluble organic carbon (A) and microbial biomass carbon (B) content of soil co-incubation with microplastics, rice straw, and both

微塑料组与对照组相比,水稻土、红壤和潮土的DOC和MBC含量增加;RS-PE与RS相比,水稻土

的DOC和MBC含量降低,红壤、潮土的DOC和MBC含量增加。由于土壤自身TOC含量的差异和

外源碳的输入,土壤类型和外源材料对DOC和MBC含量影响显著,但两者交互作用仅对DOC含量影响显著(图2)。

2.4 微塑料和秸秆添加对土壤可溶性有机碳结构的影响

紫外可见吸收光谱中获得的SUVA₂₅₄、SUVA₂₆₀和E₂/E₃可以表征可溶性有机碳的化学特性。结果显示,3种土壤中微塑料组SUVA₂₅₄和SUVA₂₆₀值略有增加,RS和RS-PE增加显著,且RS-PE处理最大(红壤除外)(表3),表明添加微塑料和水稻秸秆均增加了可溶性有机碳的芳香性和疏水性。

表3 土壤与微塑料、水稻秸秆及二者共培养后DOC紫外光谱指数

Table 3 UV spectral indices of DOC after co-incubation of soil with microplastics, rice straw and both

土壤类型 Soil type	处理 Treatment	SUVA ₂₅₄	SUVA ₂₆₀	E ₂ /E ₃
水稻土 Paddy soil	CK	3.53±0.12e	3.39±0.02e	6.95±0.02a
	PE	3.67±0.16e	3.42±0.16e	6.82±0.02a
	RS	3.99±0.01d	3.73±0.01d	6.63±0.05b
	RS-PE	4.44±0.26c	4.61±0.14a	6.53±0.01b
红壤 Red soil	CK	2.44±0.08g	2.25±0.08g	6.62±0.04b
	PE	2.82±0.09f	2.60±0.07f	6.36±0.06c
	RS	2.81±0.04f	2.60±0.04f	6.31±0.07c
	RS-PE	2.60±0.04fg	2.42±0.05fg	6.10±0.06d
潮土 Fluvo-aquic soil	CK	4.41±0.21c	4.12±0.21c	6.29±0.12c
	PE	4.67±0.12bc	4.35±0.10b	5.78±0.09e
	RS	4.86±0.15ab	4.53±0.16ab	5.62±0.03f
	RS-PE	5.03±0.10a	4.72±0.10a	5.18±0.16g

各土壤中E₂/E₃值大小顺序为水稻土>红壤>潮土,不同土壤的DOC相对分子质量差异显著($P<0.001$)。但各土壤中添加外源材料后,E₂/E₃值的大小为CK>PE>RS>RS-PE,表明外源材料添加后DOC相对分子质量增加,且RS-PE处理的DOC相对分子质量最大。所以,微塑料和水稻秸秆的添加有助于大分子、高芳香性和疏水性DOC组分的产生和积累(表3)。3种土壤中SUVA₂₅₄与SUVA₂₆₀呈显著正相关,而与E₂/E₃呈负相关(表4)。

三维荧光光谱-平行因子分析被用于识别和表征DOC的组分。如图3所示,在土壤样品中鉴定出3种腐殖质类化合物(C1、C2、C3)^[22]。组分C1的特征峰位于Ex(excitation wavelength) 280 nm、Em(emission wavelength) 420 nm处,其相对分子质量较大,相对稳定^[23];组分2的特征峰位于Ex 275 nm、Em

475 nm处,主要为类富里酸和胡敏酸,其结构中有高相对分子质量和高芳香度基团的存在^[24];组分C3的特征峰位于Ex 325 nm、Em 385 nm处,荧光特性简单,易被氧化分解^[25]。

DOC的3种荧光组分的荧光强度显示,类富里酸(C1、C3)和类胡敏酸(C2)分别占69.92%~76.06%和23.94%~30.08%,表明类富里酸组分是供试土壤DOC中的主要物质。3种土壤中相同处理下各组分荧光强度大小为潮土>水稻土>红壤(图4)。与CK相比,仅添加微塑料或水稻秸秆的处理,在潮土中3个组分的含量均显著增加,红壤中显著增加C2和C3组分,而水稻土中C1和C2组分无显著变化。同时添加微塑料和水稻秸秆,在水稻土和潮土中RS-PE组荧光强度最高且增加显著,而红壤中不显著。土壤中添加微塑料和水稻秸秆均会使C1、C2、C3组分荧光强度增加,且土壤类型和外源材料添加及两者交互作用对DOC各组分影响显著(图4)。

2.5 土壤CO₂累积释放量与DOC、MBC含量及紫外光谱指数的关系

土壤CO₂累积释放量与DOC、MBC含量及紫外光谱指数的相关系数见表4,3种土壤CO₂累积释放量均与DOC和MBC含量呈显著正相关,与E₂/E₃呈显著负相关,表明微生物易利用小分子的可溶性有机碳,而土壤中留存的DOC相对分子质量较大。水稻土和潮土中CO₂累积释放量与SUVA₂₅₄和SUVA₂₆₀呈显著正相关,红壤中则不显著。水稻土和潮土中DOC含量与SUVA₂₅₄和SUVA₂₆₀呈显著正相关,与E₂/E₃呈显著负相关,表明水稻土和潮土的DOC主要由大分子的高芳香性和疏水性组分构成。

3 讨论

3.1 添加微塑料和水稻秸秆对土壤性质及活性有机碳的影响

本研究结果表明,微塑料进入土壤后显著增加TOC含量和C/N比;添加水稻秸秆后,潮土和水稻土的TOC含量均显著增加。土壤类型和添加外源材料及两者交互作用对TOC和C/N比影响显著(表2)。微塑料和水稻秸秆有机碳含量较高(表1),添加后增加了土壤TOC和C/N。与CK相比,微塑料增加了3种土壤的DOC和MBC含量,与Gao等^[26]研究结果一致;RS-PE与RS相比,红壤、潮土的DOC和

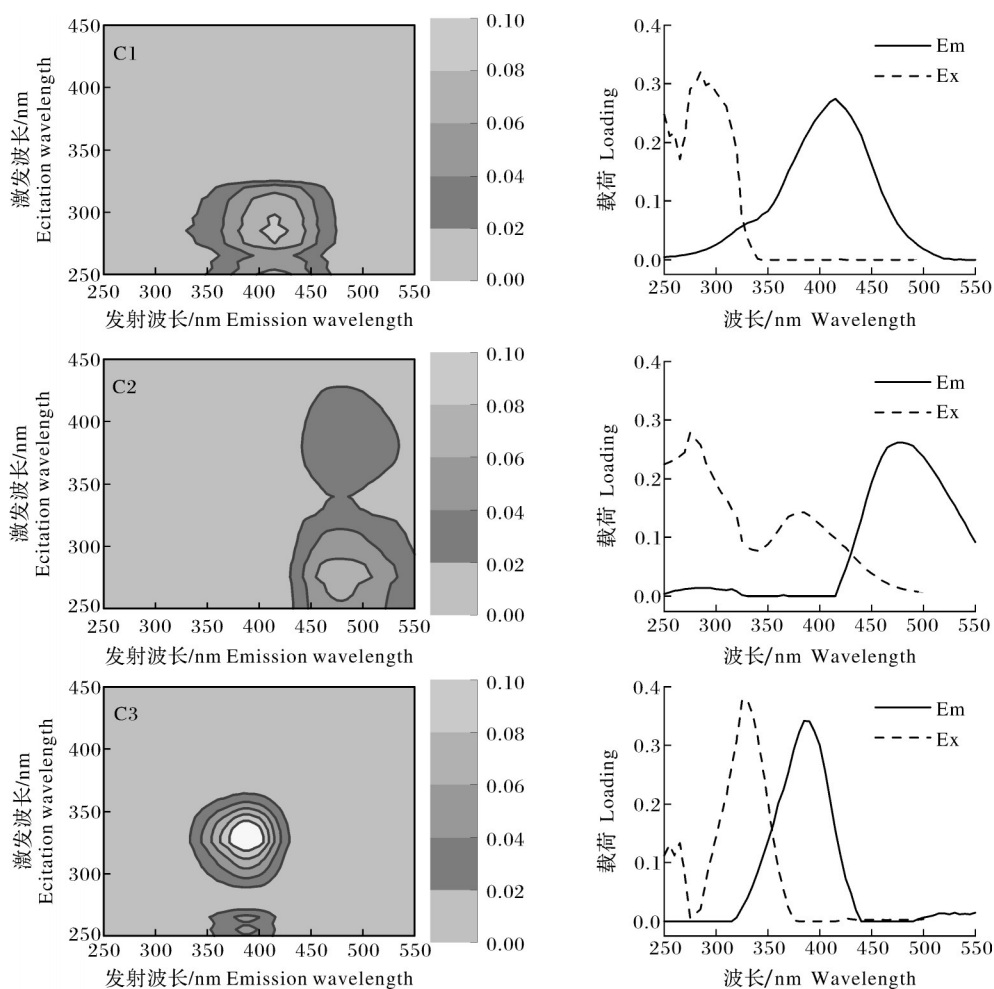


图3 三维荧光光谱-平行因子分析确定的3种组分(C1、C2、C3)的荧光光谱及载荷
Fig. 3 Fluorescence spectra and loadings of the three components(C1, C2, C3)

determined by EEM -PARAFAC analysis

MBC含量增加(图2),这是因为微塑料具有强疏水性,不仅增加DOC组分芳香性和疏水性(表3),还改变土壤容重、持水能力和通气性^[9, 25],进而影响土壤中微生物的活性,这与Yu等^[27]的研究相反,他们发现,在掺入秸秆的条件下,微塑料的存在降低了MBC含量,可能是因为他们试验中微塑料含量为10%,浓度较高,起到了抑制作用。

与微塑料相比,水稻秸秆具有更多易分解的半纤维素和纤维素,相比于CK,土壤中DOC和MBC的增加量在添加秸秆组比添加微塑料组更大(图2),表明秸秆作为碳源,比微塑料给微生物生长活动提供能量多,可加速微生物繁殖^[28]。

各土壤中添加微塑料和水稻秸秆后,DOC相对分子质量均增大,且RS-PE处理的DOC相对分子质量最大,表明水稻秸秆的分解与微塑料存在一定的拮抗作用,能缓解微塑料的毒害,维持土壤

健康^[29-30]。

3.2 添加微塑料和水稻秸秆对土壤CO₂累积矿化量的影响

有机碳矿化是碳循环的重要部分,本试验中培养初期(0~30 d),各处理的CO₂释放速率较快,随着培养时间延长逐渐减缓,在培养后期趋于平稳。培养初期土壤中存在较多活性有机碳和营养物质,且添加的外源材料中易分解组分较多,微生物活性强且增殖快,有机碳分解速率较快;随培养时间延长,土壤中活性有机碳及易分解组分含量降低,微生物增殖减弱,矿化速率逐渐减弱。

本试验中土壤类型、添加材料和两者交互作用对70 d内CO₂累积释放量影响显著(图1 D)。除RS处理外,3种土壤中PE和RS-PE处理下70 d内CO₂累积释放量大小为潮土>水稻土>红壤,与培养结束后pH变化趋势一致。表明土壤性质(土壤TOC

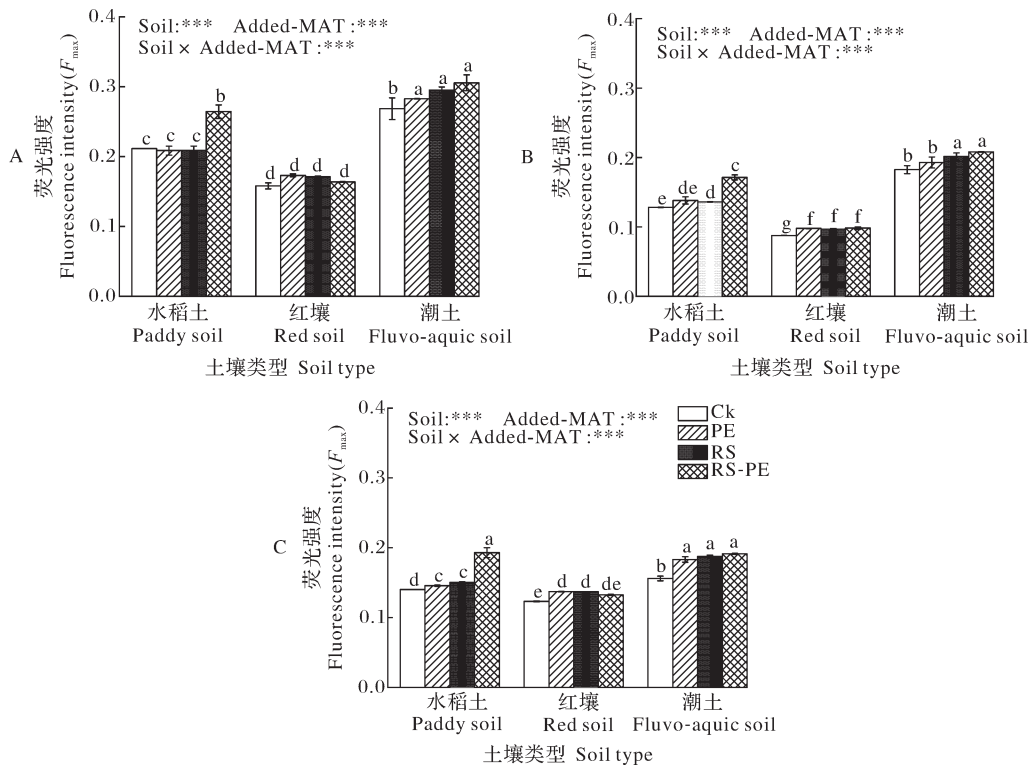


图4 土壤与微塑料、水稻秸秆及二者共培养后DOC组分C1(A)、C2(B)和C3(C)的荧光强度

Fig. 4 Fluorescence intensity of DOC components C1(A), C2(B) and C3(C) after co-incubation of soil with microplastics, rice straw, and both

表4 CO₂累积释放量与DOC、MBC含量及紫外光谱指数的相关系数

Table 4 Correlation coefficients of cumulative CO₂ emission with DOC, MBC and UV spectral index

土壤类型 Soil type	指标 Indicator	DOC	MBC	SUVA ₂₅₄	SUVA ₂₆₀	E ₂ /E ₃
水稻土 Paddy soil	CO ₂ 累积释放量 Cumulative CO ₂ emission	0.956**	0.685*	0.712**	0.599*	-0.867**
	DOC		0.684*	0.842**	0.688*	-0.905**
	MBC			0.445	0.466	-0.683*
	SUVA ₂₅₄				0.854**	-0.861**
	SUVA ₂₆₀					-0.832**
红壤 Red soil	CO ₂ 累积释放量 Cumulative CO ₂ emission	0.907**	0.660*	0.107	0.153	-0.879**
	DOC		0.743**	0.303	0.343	-0.771**
	MBC			0.419	0.456	-0.692*
	SUVA ₂₅₄				0.998**	-0.385
	SUVA ₂₆₀					-0.419
潮土 Fluvo-aquic soil	CO ₂ 累积释放量 Cumulative CO ₂ emission	0.868**	0.722**	0.845**	0.842**	-0.961**
	DOC		0.554	0.643*	0.641*	-0.784**
	MBC			0.545	0.557	-0.756**
	SUVA ₂₅₄				0.998**	-0.846**
	SUVA ₂₆₀					-0.845**

注：*表示 $P < 0.05$ ；**表示 $P < 0.01$ 。Note: * indicates $P < 0.05$ ；** indicates $P < 0.01$ 。

含量、pH、C/N)和外源C输入均影响CO₂累积释放量^[31-32]。微塑料和水稻秸秆对CO₂累积释放量影响有差异,添加秸秆会增加土壤TOC和有效养分,为微生物生长繁殖提供丰富的基质和良好的环境,促进有机碳的矿化^[33];添加微塑料可能增加土壤碳含

量,增加土壤孔隙度,增强土壤透气性,降低土壤容重,加快土壤有机质矿化^[9]。

有机碳矿化是土壤中活性有机碳在微生物作用下分解释放CO₂的过程,土壤DOC含量直接影响土壤微生物活性,从而影响CO₂的排放。本试验中3种

土壤CO₂累积释放量均与DOC和MBC含量呈显著正相关(表4),即DOC和MBC含量增加有利于有机碳的矿化,这与郝瑞军等^[34]的研究结果一致。3种土壤CO₂累积释放量均与E₂/E₃呈显著负相关,水稻土和潮土中CO₂累积释放量与SUVA₂₅₄和SUVA₂₆₀呈显著正相关,且水稻土和潮土中DOC含量与SUVA₂₅₄和SUVA₂₆₀呈显著正相关,表明水稻土和潮土有机碳矿化同时受DOC含量和化学组成的影响,微生物优先分解相对分子质量小、低芳香性和疏水性的简单有机碳而产生CO₂。

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参考文献 References

- [1] STOCKMANN U, ADAMS M A, CRAWFORD J W, et al. The knowns, known unknowns and unknowns of sequestration of soil organic carbon[J]. *Agriculture, ecosystems & environment*, 2013, 164: 80-99.
- [2] 刘琪, 李宇虹, 李哲, 等. 不同水分条件和微生物生物量水平下水稻土有机碳矿化及其影响因素特征[J]. *环境科学*, 2021, 42(5): 2440-2448. LIU Q, LI Y H, LI Z, et al. Characteristics of paddy soil organic carbon mineralization and influencing factors under different water conditions and microbial biomass levels [J]. *Environmental science*, 2021, 42 (5) : 2440-2448 (in Chinese with English abstract).
- [3] 李忠佩, 张桃林, 陈碧云. 可溶性有机碳的含量动态及其与土壤有机碳矿化的关系[J]. *土壤学报*, 2004, 41(4): 544-552. LI Z P, ZHANG T L, CHEN B Y. Dynamics of soluble organic carbon and its relation to mineralization of soil organic carbon [J]. *Acta pedologica sinica*, 2004, 41 (4) : 544-552 (in Chinese with English abstract).
- [4] 安立会, 李欢, 王菲菲, 等. 海洋塑料垃圾污染国际治理进程与对策[J]. *环境科学研究*, 2022, 35(6): 1334-1340. AN L H, LI H, WANG F F, et al. International governance progress in marine plastic litter pollution and policy recommendations [J]. *Research of environmental sciences*, 2022, 35 (6) : 1334-1340 (in Chinese with English abstract).
- [5] 马贵, 丁家富, 周悦, 等. 固原市农田土壤微塑料的分布特征及风险评估[J]. *环境科学*, 2023, 44(9): 5055-5062. MA G, DING J F, ZHOU Y, et al. Distribution characteristics and risk assessment of microplastics in farmland soil in Guyuan [J]. *Environmental science*, 2023, 44(9): 5055-5062 (in Chinese with English abstract).
- [6] THOMPSON R C, OLSEN Y, MITCHELL R P, et al. Lost at sea: where is all the plastic? [J]. *Science*, 2004, 304 (5672): 838.
- [7] 潘伟亮, 罗玲利, 敖良根, 等. 微塑料在水土环境中的来源、危害及其检测方法评述[J]. *应用化工*, 2022, 51(5): 1508-1513. PAN W L, LUO L L, AO L G, et al. The sources, hazards and detection methods of microplastics in soil and water environment [J]. *Applied chemical industry*, 2022, 51 (5) : 1508-1513 (in Chinese with English abstract).
- [8] HUANG Y, LIU Q, JIA W Q, et al. Agricultural plastic mulching as a source of microplastics in the terrestrial environment [J/OL]. *Environmental pollution*, 2020, 260: 114096 [2024-01-16]. <https://doi.org/10.1016/j.envpol.2020.114096>.
- [9] 鄂玉联, 谭兰兰, 安梦洁, 等. 高分子化合物对盐渍化棉田土壤团聚体组成及棉花产量的影响[J]. *南方农业学报*, 2017, 48(11): 1989-1993. E Y L, TAN L L, AN M J, et al. Effects of polymer compounds on soil aggregate composition and cotton yield in salted cotton field [J]. *Journal of southern agriculture*, 2017, 48 (11) : 1989-1993 (in Chinese with English abstract).
- [10] 李晶晶, 白岗栓. 聚丙烯酰胺的水土保持机制及研究进展[J]. *中国水土保持科学*, 2011, 9(5): 115-120. LI J J, BAI G S. Mechanism of PAM on soil and water conservation and its development [J]. *Science of soil and water conservation*, 2011, 9 (5) : 115-120 (in Chinese with English abstract).
- [11] 费禹凡, 黄顺寅, 王佳青, 等. 设施农业土壤微塑料污染及其对细菌群落多样性的影响[J]. *科学通报*, 2021, 66(13): 1592-1601. FEI Y F, HUANG S Y, WANG J Q, et al. Microplastics contamination in the protected agricultural soils and its effects on bacterial community diversity [J]. *Chinese science bulletin*, 2021, 66 (13) : 1592-1601 (in Chinese with English abstract).
- [12] LIU M L, FENG J G, SHEN Y W, et al. Microplastics effects on soil biota are dependent on their properties: a Meta-analysis [J/OL]. *Soil biology and biochemistry*, 2023, 178: 108940 [2024-01-16]. <https://doi.org/10.1016/j.soilbio.2023.108940>.
- [13] HUANG W, WU J F, PAN X H, et al. Effects of long-term straw return on soil organic carbon fractions and enzyme activities in a double-cropped rice paddy in South China [J]. *Journal of integrative agriculture*, 2021, 20(1): 236-247.
- [14] WANG Y L, WU P N, MEI F J, et al. Does continuous straw returning keep China farmland soil organic carbon continued increase: a meta-analysis [J/OL]. *Journal of environmental management*, 2021, 288: 112391 [2024-01-16]. <https://doi.org/10.1016/j.jenvman.2021.112391>.
- [15] 李新华, 郭洪海, 朱振林, 等. 不同秸秆还田模式对土壤有机碳及其活性组分的影响[J]. *农业工程学报*, 2016, 32(9): 130-135. LI X H, GUO H H, ZHU Z L, et al. Effects of different straw return modes on contents of soil organic carbon and fractions of soil active carbon [J]. *Transactions of CSAE*, 2016, 32(9): 130-135 (in Chinese with English abstract).

- [16] 鲁如坤.土壤农业化学分析方法[M].北京:中国农业科学技术出版社,2000.LU R K.Methods of soil agrochemical analysis [M]. Beijing: China Agricultural Science and Technology Press,2000(in Chinese).
- [17] XU P D, ZHU J, WANG H, et al.Regulation of soil aggregate size under different fertilizations on dissolved organic matter, cellobiose hydrolyzing microbial community and their roles in organic matter mineralization[J/OL].Science of the total environment, 2021, 755: 142595 [2024-01-16]. <https://doi.org/10.1016/j.scitotenv.2020.14259>.
- [18] LEE H S, HUR J, LEE M H, et al.Photochemical release of dissolved organic matter from particulate organic matter: spectroscopic characteristics and disinfection by-product formation potential[J].Chemosphere, 2019, 235: 586-595.
- [19] MURPHY K R, STEDMON C A, GRAEBER D, et al.Fluorescence spectroscopy and multi-way techniques. PARAFAC [J].Analytical methods, 2013, 5(23): 6557-6566.
- [20] BAHRAM M, BRO R, STEDMON C, et al.Handling of Rayleigh and Raman scatter for PARAFAC modeling of fluorescence data using interpolation [J]. Journal of chemometrics, 2006, 20(3/4): 99-105.
- [21] STEDMON C A, BRO R. Characterizing dissolved organic matter fluorescence with parallel factor analysis: a tutorial[J]. Limnology and oceanography: methods, 2008, 6(11): 572-579.
- [22] LIN H, GUO L D. Variations in colloidal DOM composition with molecular weight within individual water samples as characterized by flow field-flow fractionation and EEM-PARAFAC analysis[J]. Environmental science & technology, 2020, 54(3): 1657-1667.
- [23] 李帅东, 张明礼, 杨浩, 等. 昆明松华坝库区表层土壤溶解性有机质(DOM)的光谱特性[J]. 光谱学与光谱分析, 2017, 37(4): 1183-1188. LI S D, ZHANG M L, YANG H, et al. Spectroscopic characteristics of dissolved organic matter from top soils on Songhuaba Reservoir in kunming [J]. Spectroscopy and spectral analysis, 2017, 37(4): 1183-1188 (in Chinese with English abstract).
- [24] 郭卫东, 黄建平, 洪华生, 等. 河口区溶解有机物三维荧光光谱的平行因子分析及其示踪特性[J]. 环境科学, 2010, 31(6): 1419-1427. GUO W D, HUANG J P, HONG H S, et al. Resolving excitation emission matrix spectroscopy of estuarine CDOM with parallel factor analysis and its application in organic pollution monitoring[J]. Environmental science, 2010, 31(6): 1419-1427 (in Chinese with English abstract).
- [25] DE SOUZA MACHADO A A, LAU C W, TILL J, et al. Impacts of microplastics on the soil biophysical environment [J]. Environmental science & technology, 2018, 52(17): 9656-9665.
- [26] GAO B, YAO H Y, LI Y Y, et al. Microplastic addition alters the microbial community structure and stimulates soil carbon dioxide emissions in vegetable-growing soil [J]. Environmental toxicology and chemistry, 2021, 40(2): 352-365.
- [27] YU H, ZHANG Z, ZHANG Y, et al. Effects of microplastics on soil organic carbon and greenhouse gas emissions in the context of straw incorporation: a comparison with different types of soil [J/OL]. Environmental pollution, 2021, 288: 117733 [2024-01-16]. <https://doi.org/10.1016/j.envpol.2021.117733>.
- [28] 王季斐, 童瑶瑶, 祝贞科, 等. 不同水平外源碳在稻田土壤中转化与分配的微生物响应特征[J]. 环境科学, 2019, 40(2): 970-977. WANG J F, TONG Y Y, ZHU Z K, et al. Transformation and distribution of soil organic carbon and the microbial characteristics in response to different exogenous carbon input levels in paddy soil [J]. Environmental science, 2019, 40(2): 970-977 (in Chinese with English abstract).
- [29] 冯雪莹, 孙玉焕, 张书武, 等. 微塑料对土壤-植物系统的生态效应[J]. 土壤学报, 2021, 58(2): 299-313. FENG X Y, SUN Y H, ZHANG S W, et al. Ecological effects of microplastics on soil-plant systems [J]. Acta pedologica sinica, 2021, 58(2): 299-313 (in Chinese with English abstract).
- [30] 朱永官, 朱冬, 许通, 等. (微)塑料污染对土壤生态系统的影响: 进展与思考[J]. 农业环境科学学报, 2019, 38(1): 1-6. ZHU Y G, ZHU D, XU T, et al. Impacts of (micro) plastics on soil ecosystem: progress and perspective [J]. Journal of agro-environment science, 2019, 38(1): 1-6 (in Chinese with English abstract).
- [31] 刘志伟, 朱孟涛, 郭文杰, 等. 秸秆直接还田与炭化还田下土壤有机碳稳定性和温室气体排放潜力的对比研究[J]. 土壤通报, 2017, 48(6): 1371-1378. LIU Z W, ZHU M T, GUO W J, et al. Comparison of soil organic carbon stability and greenhouse gas emissions potential under straw or straw-derived biochar amendment [J]. Chinese journal of soil science, 2017, 48(6): 1371-1378 (in Chinese with English abstract).
- [32] 张秀玲, 鄢紫薇, 王峰, 等. 微塑料添加对橘园土壤有机碳矿化的影响[J]. 环境科学, 2021, 42(9): 4558-4565. ZHANG X L, YAN Z W, WANG F, et al. Effects of microplastics addition on soil organic carbon mineralization in *Citrus* orchard [J]. Environmental science, 2021, 42(9): 4558-4565 (in Chinese with English abstract).
- [33] 高燕, 张延, 郭亚飞, 等. 不同秸秆还田模式对土壤有机碳周转的影响[J]. 土壤与作物, 2019, 8(1): 93-101. GAO Y, ZHANG Y, GUO Y F, et al. Effect of residue return patterns on soil organic carbon turnover: a review [J]. Soils and crops, 2019, 8(1): 93-101 (in Chinese with English abstract).
- [34] 郝瑞军, 李忠佩, 车玉萍. 苏南水稻土有机碳矿化特征及其与活性有机碳组分的关系[J]. 长江流域资源与环境, 2010, 19(9): 1069-1074. HAO R J, LI Z P, CHE Y P. Characteristic of soil organic carbon mineralization and its relationship with active organic carbons in paddy soils of southern Jiangsu Province [J]. Resources and environment in the Yangtze Basin, 2010, 19(9): 1069-1074 (in Chinese with English abstract).

Effects of adding microplastics and rice straw on emission of CO₂ from typical farmland soils

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Abstract Microplastics are widely present in soil, and the problem of microplastic pollution in soil has been considered during recent years. Three typical farmland soils including the paddy soil, the red soil and the fluvo-aquic soil were used to conduct indoor incubation experiments for 70 days to study the effects of microplastic pollution on the mineralization of organic carbon in soil under rice straw returning. 4 treatments including the control (CK), addition of microplastics (polyethylene, PE), addition of rice straw (RS), addition of microplastics and rice straw (RS-PE) were set for each type of soil. The physical and chemical properties, the emission of CO₂, the content of soluble organic carbon (SOC) and microbial biomass carbon (MBC) in each soil treated were measured. The results showed that the cumulative emission of CO₂ from three types of soils within 70 days under the PE and RS-PE treatment was in the order of fluvo-aquic soil > paddy soil > red soil. Compared with CK, the cumulative emission of CO₂ from three types of soils significantly increased after RS treatment, and the cumulative emission of CO₂ from the red soil and fluvo-aquic soil significantly increased after PE treatment. Adding PE and RS increased the content of SOC and MBC in three types of soils, promoting the mineralization of organic carbon in soil. The E₂/E₃ value in three types of soils was in the order of paddy soil > red soil > fluvo-aquic soil. The addition of PE and RS increased the average relative molecular weight, aromaticity, and hydrophobicity of SOC in soil. The mineralization of organic carbon in three types of soils was significantly and negatively correlated with the average relative molecular weight of SOC (E₂/E₃), and the mineralization of organic carbon in paddy and fluvo-aquic soil was significantly and positively correlated with the aromaticity (SUVA₂₅₄) and hydrophobicity (SUVA₂₆₀) of SOC. It is indicated that the addition of microplastics and rice straw significantly affect the mineralization of organic carbon in soil.

Keywords microplastics; soil; rice straw; mineralization of organic carbon; emission of carbon dioxide

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